



## **EQUILIRIUM, KINETICS AND THERMODYNAMIC STUDY OF Pb(II) IONS FROM AQUEOUS SOLUTION BY USING FIRED CERAMIC**

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### **Abstract:**

*Toxic heavy metals are some of the most common hazardous pollutants found in industrial wastewater. The industrial activities and technological development lead to a discharge of significant amount of heavy metal ions including Pb(II) in their effluent which influence the health of human population and also affect aquatic life. The long term exposure of Pb(II) cause decrease in performance of nervous system, weakness in bone joints, increase in blood pressure, anaemia. High level exposure of lead can severely damage brain, kidneys and may cause miscarriage in pregnant women. Therefore, it is a matter of great concern to remove lead from industrial water before discharging it into the water bodies. The aim of the present study is to explore the feasibility of using Fired Ceramic as adsorbent for the removal of Lead ions from aqueous solutions under different conditions. Batch studies have been conducted to investigate the influence of various parameters of adsorption such as pH, contact time, adsorbent dose and metal ions concentration on the adsorption of lead by fired ceramic. It has been observed that adsorption percentage of Pb(II) increases to 96.320% with increase in pH up to pH 4 and decreases with increase of metal ion concentration. Equilibrium time required for the adsorption of lead ions by fired ceramic was found to be 2 hours. The kinetic studies revealed that adsorption of lead by fired ceramic follow both first and second order kinetics. The highest value of  $R^2 = 0.997$  for Langmuir isotherm indicates adsorption of Pb(II) by fired ceramic obey Langmuir isotherm more appropriately than Freundlich and Temkin model. These results suggested that the fired ceramic can be effectively employed as cost free and environment friendly adsorbent for the removal of heavy metals in aqueous system.*

**Key Words:** Fired Ceramic, Adsorption, Batch Process, Langmuir, Freundlich, Temkin Isotherm, Pseudo-First Order, Second Order & Intra-Particle Diffusion.

### **1. Introduction:**

For decades, urban and industrial pollution has led to a gradual environmental degradation of the natural environment (atmosphere, water, soil). This phenomena is serious and requires great efforts to correct the situation and prevent its proliferation. The fast paced development of industries such as metal mining operations, fertilizers and paper industries and pesticides have deliberately discharge various types pollutants into the environments especially in developing countries. Among other issues, water contaminations by heavy metals are more pronounced than other pollutants especially when heavy metals are exposed to the natural ecosystems. Heavy metals refers to any element with the atomic weight between 63.5 and 200.6 and a specific gravity greater than 5.0 [1]. Heavy metals are non-biodegradable and tend to accumulate in human body causing various disorders [2]. Heavy metals have been expressively released into the environment due to rapid industrialisation and have created a major global concern [3]. Although the presence of heavy metals does not cause immediate danger to the population, it can lead to sustained, irreversible or wider

contamination of the environment. This in turn could lead to long term contamination of the surrounding soil and groundwater [4].

Lead (Pb) is considered as one of the priority metals from the point of view of potential health hazards to human, and it is listed by EPA as one of 129 priority pollutants [5]. Lead as a pollutant is a major concern as it has been used as one of the raw materials for battery manufacturing, printing, pigments, fuels, photographic materials and explosive manufacturing. Lead is used to produce batteries, metal products like solder and pipes and X-ray shielding devices. Lead is highly toxic metal and as a result of related health concern, its use in several products like gasoline, paints and pipe solder has been drastically reduced in recent years. Today most common source as lead exposure in the U.S are lead based paint and possibly water pipes in older homes, contaminated soil, house hold dust, drinking water, lead crystals, lead in certain cosmetics and toys and lead glazed pottery. The presence of lead in drinking water even at low concentration may cause disease such as anaemia, encephalopathy, hepatitis and nephritic syndrome. Severe lead poisoning can cause encephalopathy with permanent damage, while moderate lead poisoning result in neurobehavioral and intelligent deficit. Lead poisoning in human causes severe damage to kidneys, nervous system, reproductive system, liver brain. According to the United States Environmental protection Agency (USEPA) the maximum permissible limits in wastewater and potable water are 0.1 mg/L and 0.015 mg/L respectively for lead (II) [6].

EPA has determined that lead is a probable human carcinogen. Permissible limit of lead in drinking water by EPA, SUEPA and WHO is  $0.05 \text{ mgL}^{-1}$  and that of Bureau of Indian Standards (BIS) is  $0.1 \text{ mgL}^{-1}$ . Lead can effects every organ and system in the body. Long term exposure of adults can results in decrease performance of nervous system, weakness in fingers and wrists or ankles, increase in blood pressure and anaemia. Exposure to high lead levels can severely damage the brain and kidneys and ultimately cause death. In pregnant women, high levels of exposure to lead may cause miscarriage. High level exposure in men can damage the organ responsible for sperm production. Therefore, It becomes necessary to remove these heavy metals from these wastewater by as appropriate treatment before releasing them into the environment [7].

Many conventional techniques have been used to reduce the concentration of heavy metals present in wastewater. Chemical precipitation and filtration, chemical oxidation and reduction, electrochemical treatment, reverse osmosis, solvent extraction and evaporation all shows several disadvantages, such as high cost, incomplete metal removal, low selectivity, high energy requirements and generation of toxic slurries that are difficult to be eliminate [8]. Ion-exchange is a highly popular method and has been widely practiced in industrial wastewater treatment process, but the application such process if often restricted because it cannot guarantee the metal concentration limits required by regulatory standards, especially when the metal concentration in the effluent are below  $100 \text{ mg/L}$  [9]. Adsorption is one of the physiochemical treatment process found to be effective in removing heavy metals from aqueous solution. The choice of the adsorbent is a key point and the literature review suggested the use of new adsorbents with high adsorption capacity could remove the metal ions effectively. Among them glazed tea mugs [10], saw dust[11], fruit peel of orange[12], rice husk[13], powdered waste sludge (PWS)[14], dried activated sludge biomass[15], flyash [16], gyttja (which has multilayer coal deposits with clay and calcareous gyttja partings) [17], kaolinite supported zero valent iron mono particles [18], Mg oxide coated betonite [19], betonite polyacrylamide composites [20], humic acid immobilised betonite composite

[21], Na-bentonite [22], , sepiolite [23], mineral soil [24], perlite [25], sandy soil [26], zeolite prepared from Egyptian kaoline, chitosan coated sand [27], pottery glaze [28] have recently been reported.

In this article the technical feasibility of various clay based and low cost adsorbent for heavy metal removal from contaminated water has been reviewed [29].

Clays have been evaluated for their purifying qualities since they are thought to act as a filter and purifier of pollutants [30], [31], [55]. Given their high permeability, clays are after used as pollutants barrier for water storage sites [32]. In this regard, multiple researchers studied the effectiveness of various type of clay in removing heavy metals from aqueous solutions including natural kaolinite clay [33], natural bentonite [34], montmorillonite, sodium dodecyl sulphate modified iron pillared montmorillonite [35]. Adsorption is the most preferred method for the removal of heavy metals from aqueous solution due to its simplicity and its high effectiveness [36, 54]. Adsorbent can be considered as cheap or low cost if it is abundant in nature, requires little processing and is a by-product waste of the industry. In the same contest Fired Ceramic is a low cost adsorbent as it is easily available in bulk in local area and is a waste produced by the ceramic industry.

#### **Materials and Methods:**

The adsorbent use in the present study is the Fired Ceramic which is obtained as waste when moulded ceramic articles broken down or get deshaped during heating in the furnace. These broken or deshaped articles were collected, grinded into fine particles and sieved through 100 mesh size and used as adsorbent for Pb(II) ions uptake from aqueous solution.

To observed the surface morphology of the adsorbent before and after adsorption of Pb(II), SEM analysis was employed by using Scanning Tunnelling Microscope, JEOL-JSM 6510 LV at USIF (University Sophisticated Instrumentation Facility) Aligarh Muslim University, Aligarh. The types of binding groups present on the adsorbent (RCC) were identified by using Perkin-Elmer Fourier Transformer Spectrophotometer (FTIR) Version 10.03.08 by using KBr pellet technique in the wavelength range  $450\text{--}4000\text{ cm}^{-1}$  at Instrumentation Lab, Department of Chemistry, AMU, Aligarh.

All chemicals used in the study were of analytical grade purchased from Merck Pvt. Ltd. A stock solution of 500 mg/L of Pb(II) was prepared by dissolving required amount of Lead Nitrate,  $\text{Pb}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$  in double distilled water in a 500 ml standard flask. This stock solution was used to prepare the working solutions of desired concentrations for all experiments. pH values of the solutions was adjusted by adding 0.1 mol/L NaOH or 0.1 mol/L HCl. The concentrations of the solution before and after adsorption were analysed by using GBC-902, Atomic Absorption Spectrophotometer (AAS).

#### **Batch Studies:**

Batch process were conducted at room temperature  $27 \pm 1^\circ\text{C}$  to study the adsorption of Pb(II) ions by fired ceramic. An accurately weighed 0.5gm of fired ceramic was placed in 100ml stoppered conical flask containing 50 ml of Pb(II) ions solution of concentration 50mg/L. This solution was shaken in a rotary shaker for about 15 min and kept for 24 hours to attain equilibrium. All the operations such as effect of pH, adsorbent dose, contact time were conducted at room temperature with 50ml Pb(II) ions solutions of 50mg/L except during the study of effect of concentration where 10 to 200 mg/L Pb(II) ions concentration solutions were used. After that the mixture was filtered by using whattman filter paper 41 to separate the adsorbent from

supernatant liquid. The final concentration of metal ions was analysed by AAS. The amount of metal ion adsorbed was calculated by subtracting final concentration from initial concentration.

The effect of pH on the adsorption of Pb(II) was studied by batch process as follows: 50 ml of 50 mg/L (Initial concentration) metal solution was taken in a 100 ml conical flask. The pH of the solution was adjusted by adding 0.1 M HCl and 0.1 M NaOH solutions by using SYSTRONIC-09 pH meter. This solution was treated with 0.5 gm of adsorbent and left for attaining equilibrium. After attaining equilibrium final concentration of Pb (II) in the solution was determined by AAS.

A series of 100 ml conical flasks, each having 0.5 g fired ceramic and 50 ml solution of 50 mg/L concentration were shaken in a rotary shaker and at the predetermined intervals. The solution of the specified flask was taken out and filtered. The concentration of Pb(II) in the filtrate was determined by AAS.

A series of 100 ml conical flasks, each containing 50 ml solution of 50 mg/L concentration were treated with varying amount adsorbent (0.1 – 1.0g) and shaken in a rotary shaker. After that the solution filtered and the concentration of Pb(II) in the filtrate was determined by Asante amount of lead adsorbed in each case was calculated.

A series of 100 ml conical flasks, having 0.5 g of adsorbent and 50 ml solution each of varying concentration from 10 - 200 mg/L concentration were shaken in a rotary shaker. After predetermined interval the solution was filtered and concentration of Pb(II) in the filtrate was determined by AAS. The amount of lead adsorbed in each case was calculated.

The removal percentage (R%) of Pb(II) ions and Adsorption capacity,  $q_e$ (mg/g) adsorbent were calculated for each run by the following expression-

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$$R\% = \frac{(C_i - C_e)}{C_i} \times 100$$

$$q_e = \frac{(C_i - C_e)}{m} \times V$$

Where,

$C_i$  = Initial concentration of metal ions in the solution

$C_e$  = Final concentration of metal ions in the solution

$V$  = Volume of the solution (L)

$m$  = Mass of the adsorbent (in gm)

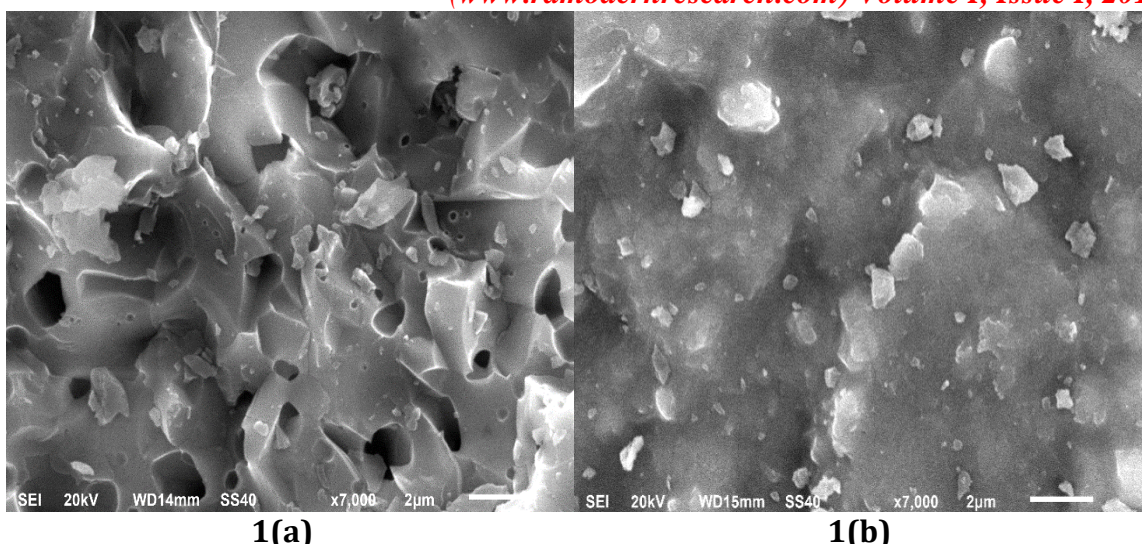
## **Results and Discussion:**

### **Characterization of Adsorbent:**

#### **SEM Analysis:**

Scanning Electron Microscope Analysis was carried out to examine the surface morphology of native and Pb(II) treated fired ceramic which are shown in fig. 1(a) and (b). The surface of fired ceramic in both cases appears to be irregular and porous and thus revealing the presence of suitable adsorbent sites. The morphology of surface is slightly changed after adsorption as before adsorption surface appeared to be less dense with light grey patches but after adsorption surface become slightly regular and denser with dark grey patches showing the adherence of Pb(II) ions on the surface.

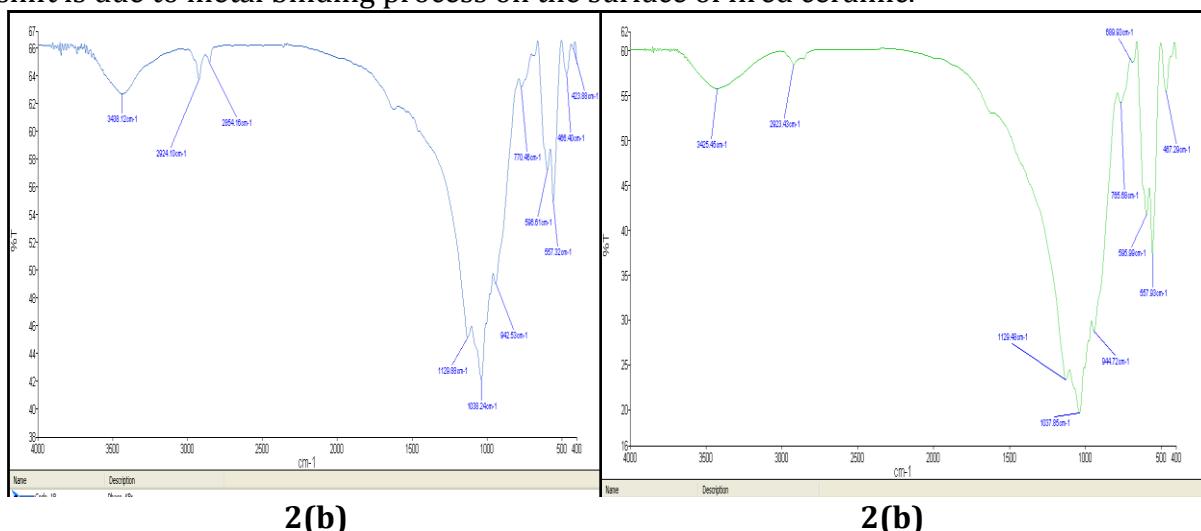




**Figure 1(a) SEM image of native fired ceramic & Figure 1(b) SEM image of Pb(II) treated fired ceramic**

### FTIR Analysis:

FTIR analysis has been done to find the type and structure of binding groups present on fired ceramic in the wavelength range  $400\text{--}4000\text{ cm}^{-1}$ . FTIR spectra of untreated and Pb(II) treated fired ceramic is shown in fig. 2(a) & (b). The FTIR spectrum showed a characteristic broad peak at  $3438.12\text{ cm}^{-1}$  due to the presence of hydrogen bonded  $\text{-OH}$  group, while small peak at  $2924.10\text{ cm}^{-1}$  and  $2854.16\text{ cm}^{-1}$  indicates the presence of C-H stretched bond in alkanes. Sharp and small peak at  $1129\text{ cm}^{-1}$  was due to the presence of C-O stretched bond. Peaks between  $900$  and  $1100\text{ cm}^{-1}$  are due to metal oxides bond. The peaks between  $500$  and  $800\text{ cm}^{-1}$  indicates the presence of C-X bonds ( $X = \text{Halogen}$ ). When FTIR spectra of untreated and Pb(II) treated adsorbent were compared, the wave number associated with  $\text{-OH}$  group was shifted showing the interaction of Pb(II) occurs at  $\text{-OH}$  groups. All these changes show that this shift is due to metal binding process on the surface of fired ceramic.

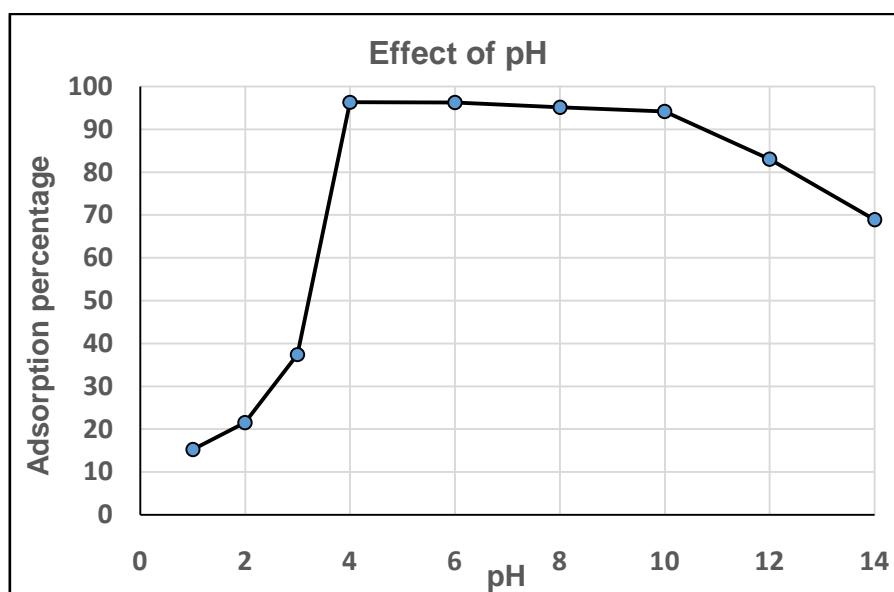
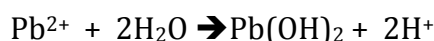
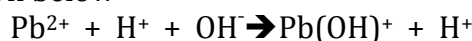


**Figure 2 (a): FTIR Image of RCC before adsorption & Figure 2 (b): FTIR Image of RCC after adsorption**

### Effect of pH:

The pH of the solution has significant effect on the uptake of Pb(II) ions. The amount of Pb(II) adsorbed at different pH values is shown in fig. 3. The amount of

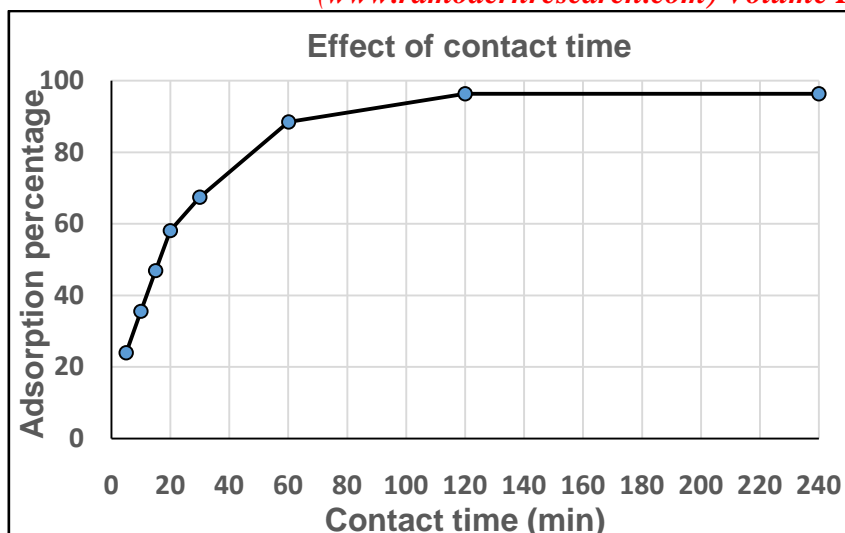
Pb(II) adsorbed on fired ceramic was found to increase from 15.2 % to 98.372% by increasing pH from 1 to 4 and then decrease slightly. The adsorption percentage of Pb(II) was least at pH 1 and increase with increasing pH, reached maximum (98.372%) at pH 4 and then decrease slowly up to pH 10. The variation in the adsorption of Pb(II) with respect to pH can be explained by considering the initial pH ( $pH_i$ ), final pH or equilibrium pH ( $pH_f$ ), and speciation of metal ions in the solution[37]. At pH 1 there is competition between  $H^+$  ions and  $Pb^{2+}$  ions for binding sites of adsorbent, therefore the adsorption of Pb(II) ions was least (15.2%). When the initial pH of the solution increases to 2 the adsorption percentage increases to 21.5%, possibly due to little less competition of  $Pb^{2+}$  ions and  $H^+$  ions. However, when initial pH raised to 4 the adsorption of Pb(II) increased up to maximum of 96.320% and will remain almost same when the initial pH adjusted to 6 (96.220%), as a result of maximum adsorption of  $Pb^{2+}$  ions along with  $H^+$  ions. Further increase in pH of the solution to 8 results slightly decrease in the adsorption of Pb(II) ions (95.120%). A similar trend was continue when pH increases to 10 (94.113%), 12 (82.99%) and 14 (68.85%) respectively. The decrease in the adsorption of Pb(II) at  $pH > 6$  can be explained on the basis of Pb(II) speciation at different pH values. The chemical species of Pb(II) formed at different pH values are  $Pb^{2+}$  at  $pH=2-4$ ,  $Pb(OH)^+$  at  $pH=4-6$ ,  $Pb(OH)_2$  at  $pH=6-10$  and  $Pb(OH)_3^-$  at  $pH=10-12$  [38, 39]. The speciation suggests that majority of the lead is adsorbed at pH 4 in the form of  $Pb^{2+}$  ions and at pH 6 lead is adsorbed in the form of lead hydroxide species due to micro precipitation as shown below



**Figure3: Effect of pH on adsorption of Pb(II) on fired ceramic**

#### **Effects of Contact Time:**

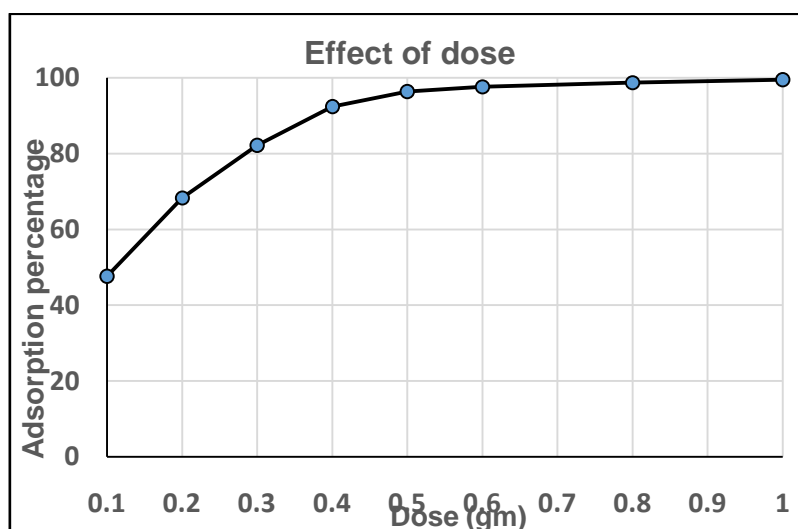
The uptake of lead ions at any particular concentration increases with contact time till the equilibrium attained as shown in fig 4. Initially the adsorption of Pb(II) was found to be fast. It may be due to more number of active sites on the adsorbent surface. After that as active sites becomes less available, the adsorption is likely to become slower and controlled process and attained equilibrium within 2 hours [40, 41].



**Figure 4: Effect of contact time on adsorption of Pb(II) on fired ceramic**

#### **Effects of Adsorbent Dose:**

The effect of dose of fired ceramic on adsorption capacity and percentage adsorption of Pb(II) is shown in fig. 5. Percent adsorption increased while adsorption capacity,  $q_e$  ( $\text{mgg}^{-1}$ ) decreased when adsorbent dose increased from 0.1 to 1.0 g. the adsorption percentage increases upto 99.45% at 1 g of adsorbent dose. This might be due to the fact that on increasing the adsorption dose, number of sites available for adsorption also increased[42]. The adsorption capacity decreases from 11.909  $\text{mg/g}$  to 2.486  $\text{mg/g}$  on increasing adsorbent dose from 0.1 g to 1.0 g. The decrease in adsorption capacity with an increasing adsorbent dose might be due to the fact that at lower adsorbent dose almost all the adsorption sites are saturated by the Pb(II) uptake but at higher adsorbent dose, the adsorption sites would be excessive for the adsorption reaction since the concentration Pb(II) as well as the volume of the solution are constant. Thus, amount of Pb(II) adsorbed per unit mass of adsorbent was decreased.[43].



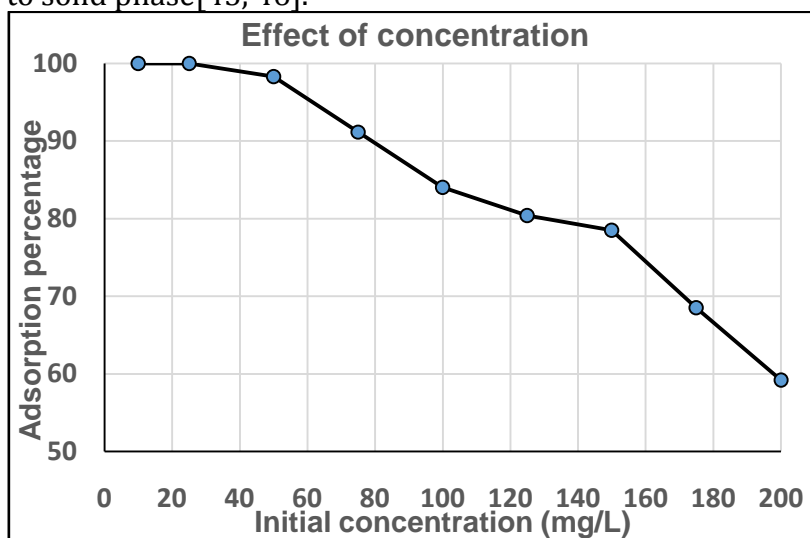
**Figure 5: Effect of adsorbent dose on adsorption of Pb(II) on fired ceramic**

#### **Effects of Initial Pb(II) ion Concentration:**

Adsorption of Pb(II) by any fired ceramic is highly dependent on the initial concentration of metal ions. The adsorption of lead was carried out at different initial Pb(II) ions concentration ranging from 10  $\text{mgL}^{-1}$  to 200  $\text{mgL}^{-1}$  and was observed that

increase of initial lead ions concentration decreases the percentage removal of lead ions from 100 % at 10 mgL<sup>-1</sup> to 59.195 % at 200 mgL<sup>-1</sup> as shown in fig. 6. It may be due to the fact that adsorbent has a limited number of active sites, which become saturated at a certain concentration [44].

It is evident from the fig. that adsorption capacity  $q_e$ , increases with increase in initial concentration of lead ions from 1 mgg<sup>-1</sup> at 10 mgL<sup>-1</sup> to 11.8 mgg<sup>-1</sup> at 200 mgL<sup>-1</sup> which may be due to fact that increased concentration gradient between the bulk solution and adsorbent surface lower down the resistance to mass transfer of Pb(II) from aqueous to solid phase[45, 46].



**Figure 6: Effect of concentration on adsorption of Pb(II) on fired ceramic**  
**Adsorption Isotherms:**

In order to indicate how molecules adsorbate is partitioned between the adsorbent and liquid phase at equilibrium as a function of adsorbate concentration, adsorption isotherms were studied. Adsorption of Pb(II) ions was carried out at concentration ranging from 10 to 200 mg/L at pH 4 for 2 hours of equilibrium time. The obtained data were analysed with the Langmuir [47], Freundlich[48] and Temkin models.

According to Langmuir model the adsorption occurs on a homogeneous surface forming monolayer of adsorbate with constant heat of adsorption for all sites without interaction between adsorbed molecules [50]. It is given by equation:

$$\frac{C_e}{q_e} = \frac{1}{b \cdot q_m} + \frac{1}{q_m} \times C_e$$

Where,  $C_e$  is the equilibrium concentration of adsorbate (mg/l),  $q_e$  is the adsorption capacity (mg/g),  $b$  and  $q_m$  are the Langmuir constants.  $q_m$  is the amount of lead required to form monolayer (mg/g), and  $b$  is the constant related to energy of adsorption.

The value of  $b$  and  $q_m$  can be calculated from the slope and intercept of the linear plot of  $C_e/q_e$  Vs  $C_e$  as shown in fig. 7. The Freundlich equation has been used for the adsorption of lead(II) on fired ceramic is represented as-

$$\text{Log} q_e = \text{log} K_f + 1/n \text{ log} C_e$$

Where,  $q_e$  is the amount of lead adsorbed,  $C_e$  is the equilibrium concentration of the lead ions in the solution (mg/L), and  $K_f$  and  $n$  are the constants for adsorption capacity and intensity of adsorption respectively. Plots of  $\text{log} q_e$  vs.  $\text{log} C_e$  has been shown in fig. 8 and the values of  $K_f$  and  $n$  was calculated from slopes and intercepts of the plot.



Temkin isotherm assumes that the decrease in the heat of adsorption is linear rather logarithmic, as implied in the Freundlich isotherm. It assumes that heat of adsorption on the surface of the adsorbent decrease linearly with coverage due to adsorbent-adsorbate interactions [49]. The linear form of Temkin equation can be represented as

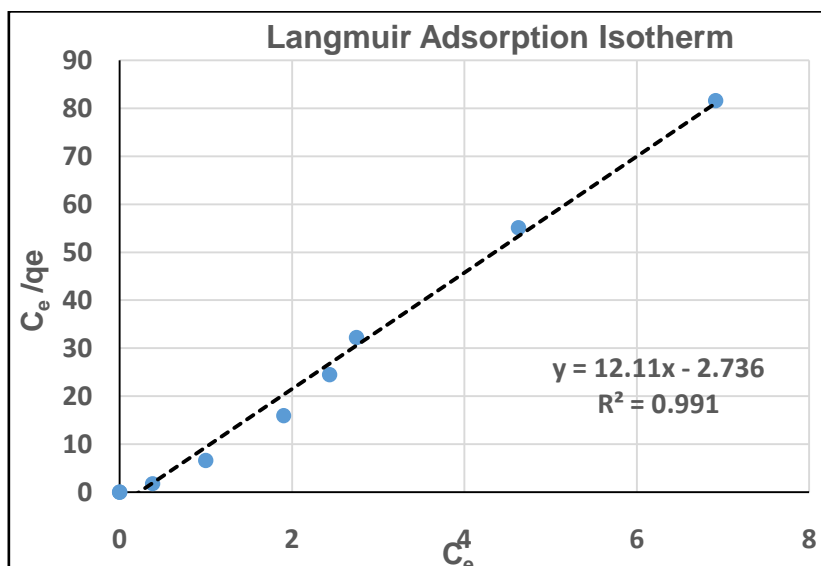
$$q_e = RT/b \times \ln A + (RT/b) \times \ln C_e$$

or

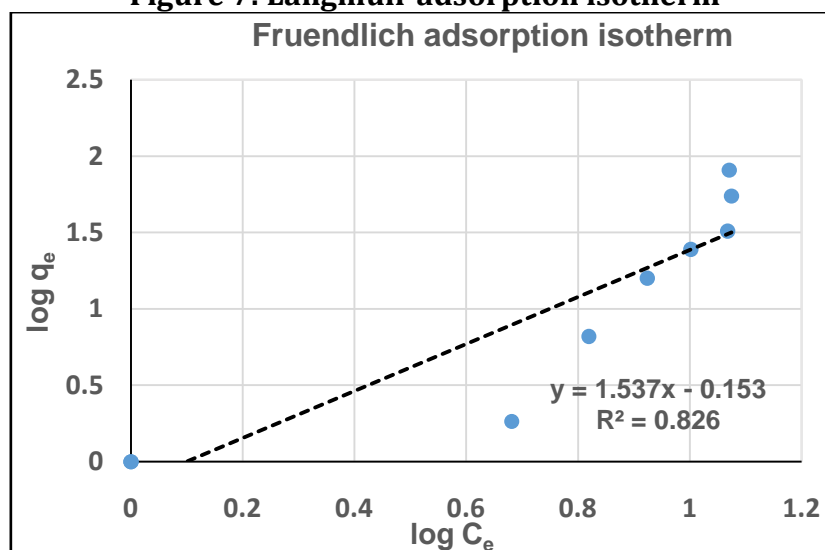
$$q_e = B \ln A + B \ln C_e$$

Where,  $B=(RT/b)$ ,  $R$  is universal gas constant,  $T$  as absolute temperature and  $b$  is a constant. The quantities  $A$  ( $\text{gL}^{-1}$ ) and  $B$  are Temkin constants related to the adsorption potential and heat of adsorption respectively. The values of  $A$  and  $B$  can be calculated from slopes and intercepts of the plot of  $q_e$  vs.  $\ln C_e$  as shown in fig. 9.

The obtained from these models and values of correlation coefficients ( $R^2$ ) 0.9917, 0.8266 and 0.9611 for the Langmuir, Freundlich and Temkin isotherms respectively indicates that adsorption of Pb(II) on fired ceramic obey best Langmuir isotherm. The obtained data also revealed the better applicability of the Temkin isotherm than Freundlich isotherm.



**Figure 7: Langmuir adsorption isotherm**



**Figure 8: Freundlich adsorption isotherm**

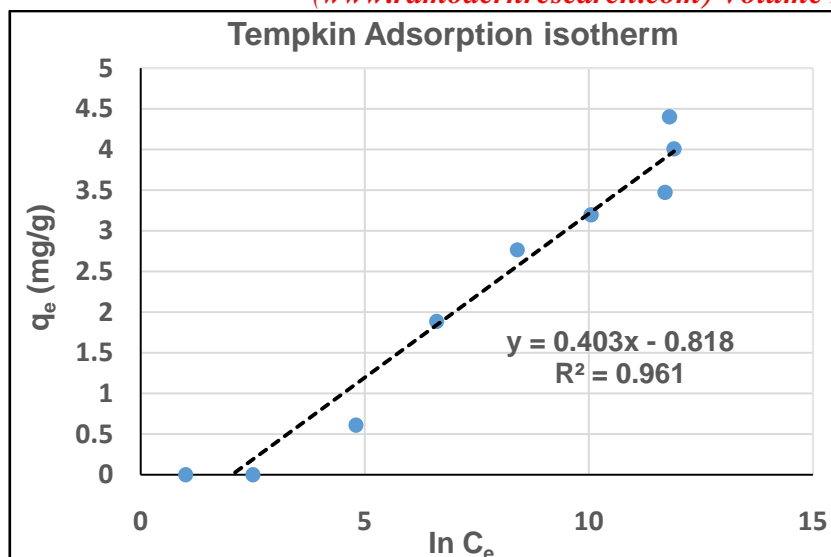


Figure 9: Tempkin adsorption isotherm

### Adsorption Kinetic Studies:

In order to investigate the mechanism of adsorption of Pb(II) on fired ceramic, the kinetic data was analysed using the Lagergren first order[50],pseudo second order[51] and Intra-particle diffusion kinetic models[52].

#### The Lagergren First Order Kinetic Model:

The Lagergren first order rate equation is represented as:

$$\ln(q_e - q_t) = \ln q_e - k_1 t$$

Or

$$\log (q_e - q_t) = \log q_e - k_1 t / 2.303$$

Where,  $q_e$  and  $q_t$  are the amounts of Pb(II) adsorbed ( $\text{mg g}^{-1}$ ) at equilibrium and at time  $t$ , respectively.  $k_1$  is the Lagergren rate constant ( $\text{min}^{-1}$ ). Plots of  $\ln(q_e - q_t)$  versus  $t$  has been shown in fig.10 and values of  $q_e$  and  $k_1$  have been calculated from the slope and intercept respectively.

#### The Pseudo-Second-Order Kinetic Model:

The adsorption data have been applied to pseudo second-order kinetic model also. The equation is represented as:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t$$

Where,  $K_2$  ( $\text{g/mg/min}$ ) is the rate constant of second order adsorption and can be obtained from plot of  $t/q_t$  against  $t$  which has been shown in fig. 11.

#### The Intra-Particle Diffusion Model:

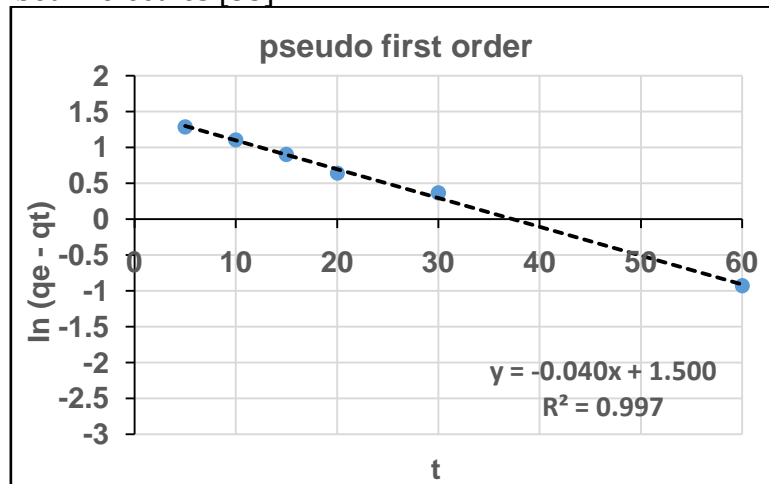
The kinetic results were analyzed by the Weber and Morris intra-particle diffusion model to explain the diffusion mechanism. The model is expressed as:

$$q_t = K_d \cdot t^{1/2} + I$$

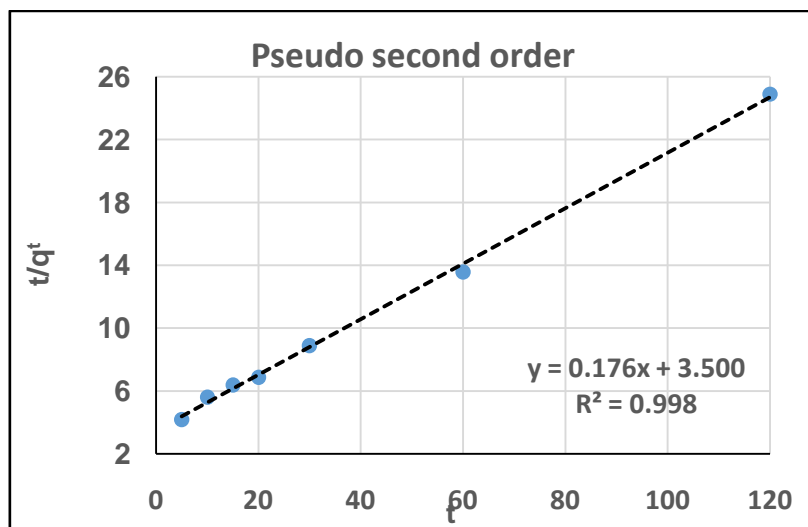
Where,  $I$  is the intercept which reflects the boundary layer effect and  $K_d$  is the intra-particle diffusion rate constant. Plot of  $q_t$  versus  $t^{1/2}$  has been shown in fig. 12. The value of  $K_d$  and  $I$  have been calculated from the slope and intercept of the plot.

All kinetic data and comparative values of correlation coefficient for pseudo first order, pseudo second order and intra particular diffusion model verifies that kinetic data is best fitted for to pseudo second order model as indicated by highest value of correlation coefficient ( $R^2 = 0.998$ ). Moreover, the high value of correlation coefficient for pseudo first order is also show its validity. Applicability of pseudo first and pseudo second order model implies that adsorption of Pb(II) on fired ceramic is a

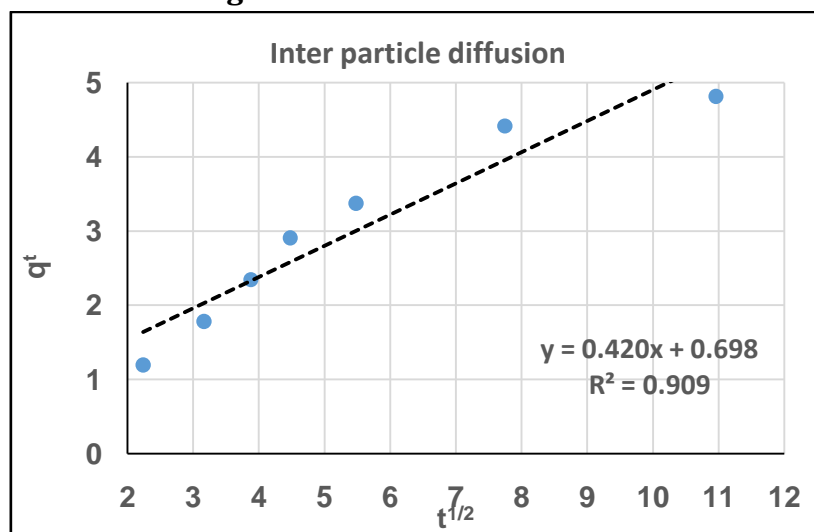
chemisorption process which involves the formation of one layer of molecules on the surface of adsorbent by chemisorption and may be followed by additional layer of physically adsorbed molecules [53].



**Figure 10: Pseudo first order model**



**Figure 11: Pseudo second order model**



**Figure12: Intra particle diffusion model**

## **Conclusion:**

The batch studies show that fired ceramic is a potential low cost adsorbent can be used to remove Pb(II) ions from aqueous solution. Adsorption of lead on fired ceramic is a fast process as equilibrium was established within 2 hrs. Adsorption of Pb(II) is a pH dependent and maximum amount of Pb(II) (96.320%) can be removed from aqueous solution at pH ranging 4-6. The experimental data is best fitted to Langmuir isotherm as indicated by the highest value of correlation coefficient ( $R^2 = 0.9917$ ). Kinetic data also suggest the applicability of both pseudo first and pseudo second order model as indicated by the values of correlation coefficients ( $R^2 = 0.9974$  and  $0.9982$ ) which also confirms that adsorption of Pb(II) on fired ceramic involve chemisorption process followed by the physical adsorption of adsorbate molecules. It may be concluded from the above results that fired ceramic can be employed as potential adsorbent for the elimination of Pb(II) ions from aqueous solution since it is cost free adsorbent which is locally available in bulk.

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